

# Canopy interactions and physical stress gradients in subtidal communities

Scott Bennett,<sup>1\*</sup> Thomas Wernberg,<sup>1</sup> Thibaut de Bettignies,<sup>1</sup> Gary A. Kendrick,<sup>1</sup> Robert J. Anderson,<sup>2,3</sup> John J. Bolton,<sup>3</sup> Kirsten L. Rodgers,<sup>4</sup> Nick T. Shears,<sup>4</sup> Jean-Charles Leclerc,<sup>5,6</sup> Laurent Lévêque,<sup>5,7</sup> Dominique Davoult,<sup>5,6</sup> and Hartvig C. Christie<sup>8</sup>

## Abstract

Species interactions are integral drivers of community structure and can change from competitive to facilitative with increasing environmental stress. In subtidal marine ecosystems, however, interactions along physical stress gradients have seldom been tested. We observed seaweed canopy interactions across depth and latitudinal gradients to test whether light and temperature stress structured interaction patterns. We also quantified interspecific and intraspecific interactions among nine subtidal canopy seaweed species across three continents to examine the general nature of interactions in subtidal systems under low consumer pressure. We reveal that positive and neutral interactions are widespread throughout global seaweed communities and the nature of interactions can change from competitive to facilitative with increasing light stress in shallow marine systems. These findings provide support for the stress gradient hypothesis within subtidal seaweed communities and highlight the importance of canopy interactions for the maintenance of subtidal marine habitats experiencing environmental stress.

## Keywords

Allee effects, competition, facilitation, kelp, seaweed, species interactions, stress gradient hypothesis.

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## INTRODUCTION

Species interactions can be important drivers of community structure and function (Kikvidze *et al.* 2005) from local to global scales (Cavieres *et al.* 2014) and have the capacity to buffer ecosystems against stressful environmental conditions that could otherwise lead to habitat loss or fragmentation (Bruno *et al.* 2003). Species interactions can change along physical and biotic stress gradients from competitive in benign conditions to become increasingly facilitative under stressful conditions, as the benefits of stress amelioration from neighbours begins to outweigh the costs of living together (SGH: Stress Gradient Hypothesis, Bertness & Callaway 1994). For terrestrial plants, the SGH has received strong empirical support within a wide range of community types across physical, resource and biotic stress gradients (He *et al.* 2013). Among subtidal marine plants (herein considered as seaweeds and seagrasses), by contrast, the nature of species interactions in response to changes to the physical environment are less well resolved (Bulleri 2009).

The physical environment has long been recognised as a strong driver of species interactions in subtidal ecosystems

(Kitching 1941; Reed & Foster 1984; Witman 1987; Dayton *et al.* 1999), however, few studies have examined whether the nature of species interactions change predictably in response to physical stress gradients (Bulleri 2009). Among the few studies that have examined physical gradients in subtidal systems, changes to abiotic conditions appear to have strong influence on the nature of interactions (Wernberg *et al.* 2010; Bulleri *et al.* 2011; Bennett & Wernberg 2014). Bulleri *et al.* (2011), for example demonstrated that the nature of interactions between tube building gastropods and seaweed was influenced by both physical (sediment deposition) and biotic (grazer density) gradients. Wernberg *et al.* (2010) and Bennett & Wernberg (2014), both demonstrated that canopy-recruit interactions can become increasingly positive or remain positive towards warmer latitudes along a latitudinal climatic gradient, although the exact role of the canopy in modifying the physical environment in these two studies remains unclear.

To date, positive species interactions in subtidal communities have been primarily reported in response to high consumer pressure (Bulleri 2009) whereby chemical defences (Hay 1986; Stachowicz & Hay 1999), mechanical defences (Anderson *et al.* 1997) and physical barriers (Bennett *et al.* 2010)

<sup>1</sup>School of Plant Biology & UWA Oceans Institute, University of Western Australia, 39 Fairway, Crawley 6009, WA, Australia

<sup>2</sup>Seaweed Research Unit, Department of Agriculture, forestry and Fisheries, Pvt Bag X2, Roggebaai 8012, South Africa

<sup>3</sup>Department of Biological Sciences and Marine Research Institute, University of Cape Town, Rondebosch 7701, South Africa

<sup>4</sup>Leigh Marine Laboratory, Institute of Marine Science, University of Auckland, PO Box 349, Warkworth 0941, New Zealand

<sup>5</sup>Sorbonne Universités, UPMC Univ Paris 6, Station Biologique, Place Georges Teissier, Roscoff 29680, France

<sup>6</sup>CNRS, Station Biologique, Place Georges Teissier, UMR 7144, Roscoff 29680, France

<sup>7</sup>CNRS, FR 2424, Station Biologique, Place Georges Teissier, Roscoff 29680, France

<sup>8</sup>Section for Marine Biology, Norwegian Institute for Water Research, Oslo, Norway

\*Correspondence and present address: Department of Environment and Agriculture, Curtin University, Bentley, WA 6102, Australia. E-mail: scott.bennett1@curtin.edu.au

reduce the incidence of predation on beneficiary species. Other important drivers of positive interactions in subtidal communities include demographic constraints, (i.e. Allee effects), whereby population fitness displays positive density dependence (Berec *et al.* 2007) and resource availability, whereby benefactors can increase the nutrient availability for beneficiary species (Peterson & Heck 2001).

Physical gradients are prevalent in the subtidal, most notably from sunlight attenuation with depth (Duarte 1991), water velocity which decreases with depth and increases with wave exposure (Siddon & Witman 2003) and temperature which usually decreases with increasing depth and latitude (Smale & Wernberg 2009). While light is often considered to be a limiting resource underwater, excessive light can become stressful for organisms (Altamirano *et al.* 2004) and therefore any modification of these conditions could promote facilitation. Within the photic zone, seaweed canopies are dominant habitat formers of benthic marine ecosystems globally and dramatically alter the physical environment for organisms living within them. Light levels are altered by all canopy types ranging from the surface canopies of the giant kelp, *Macrocystis pyrifera* (Reed & Foster 1984) to subsurface canopies of species such as *Ecklonia radiata* (Wernberg *et al.* 2005). Water velocities can also be modified by seaweed canopies, reducing the amount of drag an organism experiences by over 80% in dense canopies (Johnson 2001). Under low flow velocities, individuals can grow larger, than in high flow areas (de Bettignies *et al.* 2013a) thereby providing a mechanism for subtidal canopies to promote facilitation in high-flow environments. Temperature conditions in contrast, while critical to the distribution and abundance of seaweeds (Lüning 1984) are assumed to be largely unaffected by subtidal canopy conditions due to the high mixing of seawater (Kordas *et al.* 2011). Nevertheless, dense shallow water canopies have been observed to reduce ambient temperatures, suggesting that temperature stress amelioration could occur in high-density canopies (Critchley *et al.* 1990).

Here, we test the nature of seaweed canopy interactions in subtidal marine ecosystems in response to depth and latitudinal gradients in light and temperature stress respectively. We also examine general patterns in the nature of canopy interactions under low consumer pressure in the subtidal by observing interspecific and intraspecific interactions among nine dominant canopy formers across three continents. This study advances our understanding of how subtidal canopies function under low consumer pressure and in response to increasing abiotic stress. In doing so, we broaden the generality of the stress gradient hypothesis to incorporate subtidal light climates and reveal widespread importance of positive canopy interactions in subtidal communities.

## METHODS

### Depth experiment

Intraspecific interactions within kelp, *Ecklonia radiata*, canopies were measured along a depth gradient in Hamelin Bay, southwestern Australia (34° S, 115° E), during Austral summer, December 2013 to February 2014. Nine reefs (sites)

each separated by at least 1.8 km, were chosen along a depth gradient (3, 5, 6, 9, 11, 12, 16, 18 and 19 m depth). Reefs in the region are exposed to relatively small tidal fluctuations (< 0.5 m), meaning kelps can live on very shallow reefs and experience almost full sunlight conditions while remaining constantly submerged. At the other end of the spectrum, kelps in south-western Australia live down to 50–60 m depth where they no longer form canopies but grow as solitary individuals, presumably due the limiting light availability (Marzinelli *et al.* 2015). Reefs used in our study were in the upper third of kelp depth distribution where light is not a limiting resource (Staehr & Wernberg 2009), but could form a ‘non-resource’ stress gradient from benign deeper reefs (16–20 m) to stressful high-light shallow reefs (2–5 m). All reefs had relatively flat, ‘simple’ topography and were dominated by dense, closed *Ecklonia radiata* canopies (8–12 sporophytes m<sup>-2</sup>, following Bennett & Wernberg 2014). Within each site, 30 adult kelps were haphazardly selected and tagged around the stipe. Of those, 15 individuals were left to grow inside the unaltered canopy and 15 had all surrounding neighbours removed within a 1 m radius of the tagged kelp. Experimental setup was conducted by divers on SCUBA.

Kelp growth was measured over 2 months using the hole-punch method (de Bettignies *et al.* 2013b). Two holes were punched in the centre of the primary lamina, 5 and 10 cm above the junction between the stipe and the primary lamina. Thallus extension after the 2 months was quantified in the laboratory by measuring the distance between the two holes and the stipe-lamina-junction and subtracting the initial 10 cm distance. The kelp was then sliced into segments at 5 cm intervals along the length of the primary lamina, with each segment incorporating both primary lamina and lateral (secondary blades coming off the primary lamina) tissue. The segment with the greatest biomass in the first 30 cm above the stipe-lamina-junction was used to calculate biomass accumulation (BA, or kelp growth) as  $BA = L_{\text{ext}} \cdot W_F / 5T$  where  $L_{\text{ext}}$  is lamina extension (cm),  $W_F$  = fresh weight (g) of the heaviest segment and  $T$  is the number of growth days between punching the holes and collecting the kelp (de Bettignies *et al.* 2013b).

Differences in biomass accumulation patterns between plants with and without neighbours were converted to a Relative Interaction Index (RII, Armas *et al.* 2004) to compare the nature and intensity of interactions among sites of different depths. RII was calculated as  $RII = (BA_w - BA_o) / (BA_w + BA_o)$ , where  $BA_o$  = biomass accumulation (g.FW.d<sup>-1</sup>) without neighbours and  $BA_w$  = biomass accumulation with neighbours. The RII ranges between -1 and 1 and presents a continuous scale for competitive ( $-1 \leq x < 0$ ) and facilitative ( $0 \leq x < 1$ ) interactions that is symmetrical around zero. Patterns in RII along the depth gradient were analysed by comparing a set of polynomial models of degree = 1, 2, 3 using Akaike's Information Criterion (AIC). Linear regression (AIC = 55.5) and second order polynomial (AIC = 54.8) displayed the best relationship to the data. Linear results are presented as they provided the best fit of the data and because AIC values differ by < 2 units between the models (Anderson 2008).

Kelp tissue 'health' was measured by quantifying the cover of epiphytic algal growth on the lower and upper laterals of each kelp. Epiphyte cover was quantified by haphazardly selecting four lower and four upper laterals from each replicate kelp and photographing them. Each image was then analysed by randomly allocating points to the image and counting the presence/absence of epiphytes under 15 random points on each lateral ( $n = 120$  points per kelp) using CPCe V4.1 software (Kohler & Gill 2006). Epiphyte counts on each lateral were then used to calculate the mean percent epiphyte cover per kelp.

At the end of the experimental period the physiological acclimation of the depth and canopy treatments to experimental conditions were measured by quantifying oxygen evolution of kelps using photorespirometry chambers *in situ* under their respective experimental settings (following Rodgers *et al.* In press). Photosynthetic rates were measured at ambient light levels *in situ*, and respiration rates were measured in the dark. A detailed description of the methods can be found in Appendix S1(a).

Temperature ( $^{\circ}\text{C}$ ) and illuminance (Lux) was measured *in situ* at 15 min intervals in each site throughout the course of the experiment using data loggers (onset HOBO<sup>®</sup> data.

loggers Pendant Temp-Light, Onset Computer Corporation). Hobo pendant loggers provide light intensity measurements that accurately reflect photosynthetically active radiation (PAR) measurements taken in underwater marine environments, and enable cost effective deployment across multiple (18) experimental treatments (Long *et al.* 2012). At each site, one logger was attached to a dive weight and placed within the kelp canopy to record canopy conditions. A second logger was placed in the open adjacent to the canopy-free treatments. Daily minimum, maximum and mean temperatures for each site were extracted and averaged among days for the duration of the experiment to compare between depth and canopy treatments. For light, mean, maximum and the standard deviation of light levels between the hours of 08:00–16:00 were extracted for the first 14 days of the experiment. Fouling on the logger meant that recordings taken after this time were not representative of ambient conditions. Daily values were averaged among the first 14 days to compare among depth and canopy treatments. Variation in water velocities among sites was measured using gravitational data loggers (HOBO Pendant G, Onset Computer Corporation, Bourne, MA, USA) mounted on a hemispherical float (110 mm diameter) and attached to a 750 mm length of nylon rope (12 mm thickness) which was anchored to the bottom (Evans & Abdo 2010). The accelerometer was laid flat on the float such that only 2 channels were used (x,y) to account for horizontal acceleration (*c.f.* H2O motion V2 design; Evans & Abdo 2010). The logger was set to record the acceleration ( $\text{ms}^{-2}$ ) at 2 min intervals over the duration of the experiment. At each site the relative water movement (WM) was expressed as the mean of the 95th percentile WM over the duration of the experiment.

Kelp growth, mean epiphyte cover, temperature and illuminance were compared between canopy treatments along the depth gradient using an Analysis of Covariance (ANCOVA) with 'canopy-treatment' as a categorical grouping factor (two

levels: canopy and canopy-free) and depth as a continuous covariate. Epiphyte cover data were arcsin-square-root transformed and illuminance data were log10 transformed to address the non-normality and heteroscedasticity of the data, after examining the residual vs. fitted value plot and the Q-Q plot. Growth and temperature data met assumptions of normality and homoscedasticity and no transformations took place.

### Latitudinal experiment

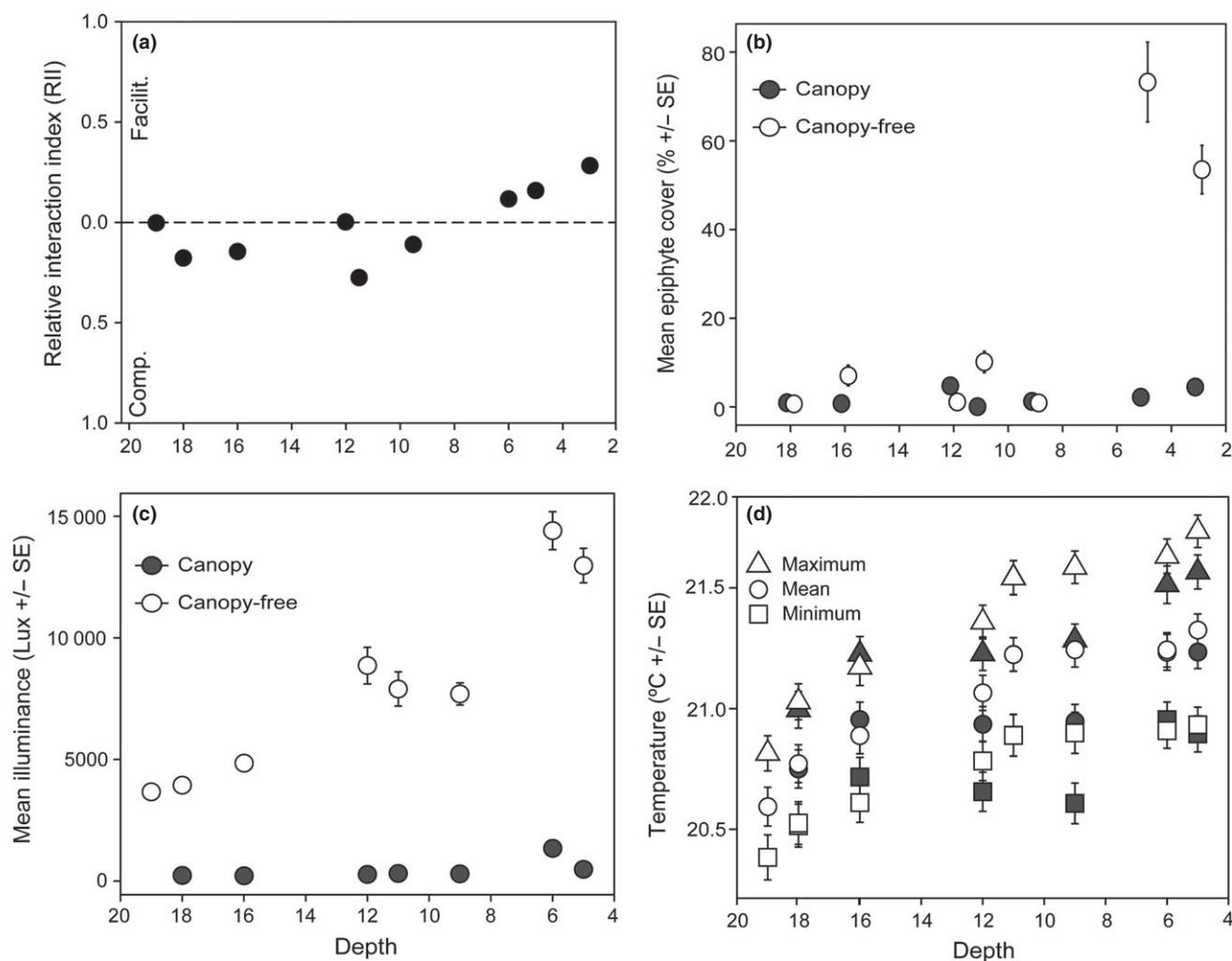
Latitudinal comparisons of intraspecific interactions took place along a *c.* 2  $^{\circ}\text{C}$  temperature gradient in south-western Australia, on two occasions during Austral spring (August–October 2012) and summer (January–March 2013), representing the periods of highest kelp growth rates and temperature stress respectively (de Bettignies *et al.* 2013b). Interaction experiments were conducted at nine sites (9–12 m depth) across three locations, Hamelin Bay (34 $^{\circ}\text{S}$ ), Marmion (32 $^{\circ}\text{S}$ ) and Jurien Bay (30 $^{\circ}\text{S}$ ), spanning the warmest third of *E. radiata*'s latitudinal distribution in Australia (*c.* 43 $^{\circ}\text{S}$ –29 $^{\circ}\text{S}$ ). This north-south trending coastline is characterised by the warm poleward flowing Leeuwin Current, consistent limestone habitat, low nutrient conditions and a stable ocean climate gradient (Smale & Wernberg 2009). Within each site, kelp growth, RII and environmental measurements in canopy and canopy-free treatments were recorded as described for the depth experiment. Tissue health was quantified for the summer sampling period only, by estimating the percentage cover on four replicate laterals from each of the lower, middle and upper sections of each kelp. Patterns in RII between spring and summer, across the latitudinal gradient were examined with an ANCOVA, using time as a fixed factor (two levels: spring and summer) and latitude as a continuous covariate. ANCOVA assumptions were checked as described for the depth experiment and no transformation took place. Mean epiphyte cover on kelps, temperature and illuminance was compared between canopy treatments along the latitudinal gradient using an ANCOVA with 'canopy-treatment' as a categorical grouping factor (two levels: canopy and canopy-free) and latitude as a continuous covariate. Epiphyte cover data were arcsin-square-root transformed and illuminance data were square-root transformed to address the non-normality and heteroscedasticity of the data.

### Interactions among different canopy types

The nature of interspecific and intraspecific interactions among different subtidal canopy species experiencing low consumer pressure were tested for nine dominant canopy species (four kelps and five fucoids) from Australia, South Africa, Norway and France (Table S1). Sites varied between regions ranging from shallow (3–4 m) tropical *Sargassum* beds in Australia, to deeper (10–12 m) cool temperate kelp forests in Norway (Table S1). Sites were selected that had dense closed seaweed canopies (Table S2) and low apparent herbivore consumer pressure. Herbivore pressure was measured during the canopy growth experiments by comparing consumption rates between canopy and canopy-free treatments (kelp) and caged

controls (*Sargassum* only, Appendix S1b, Fig. S1). Intraspecific interaction strength was tested by haphazardly selecting and tagging 30 plants, of which 15 individuals were left to grow inside the canopy and 15 had the surrounding canopy removed, by clearing a circular area around the tagged seaweed as described for the depth experiment. Clearing sizes varied in accordance with the size of the seaweed, ensuring that no contact or shading by the canopy occurred. Interspecific interactions using *Laminaria pallida* and *Sargassum decurrens* as beneficiary species were tested using the same methodology described for intraspecific interactions. Interspecific interactions using *Ecklonia radiata*, *Scytothalia dorycarpa*, *Arthrophyucus* sp. and temperate *Sargassum* sp. as beneficiaries were conducted by translocating individuals into the desired canopy and canopy-free treatments due to low natural densities at the study sites. All translocations occurred within the same site and depth contour from which they were collected,

without removing the seaweed from the water (Appendix S1c). For both interspecific and intraspecific interactions kelp (*Ecklonia radiata*, *E. maxima*, *Laminaria pallida* and *L. hyperborea*), growth rates were measured using the hole-punch method and converted to RII as described for the depth experiment. For the five fucoid species (temperate *Scytothalia dorycarpa*, *Arthrophyucus* sp and *Sargassum* sp, tropical *S. illifolium* and *S. decurrens*) net growth rates were determined by measuring the length (L) and circumference (C) of tagged individuals at the beginning and end of the experiment, and converting the measurement to fresh weight ( $W_F$ ) using the equation  $W_F = LC^2$  (following Aberg 1990). Each individual was also collected at the end of the experiment and weighed in the laboratory to the nearest 0.1 g to calibrate the biomass conversion model for each species (Fig. S2). A detailed description of the site and species specific methods can be found in Appendix S1c.



**Figure 1** Nature of intraspecific interactions of *Ecklonia radiata* along a subtidal depth gradient from 3 to 19 m. (a) Relative Interaction Index (RII) of *E. radiata* based on mean growth rates in canopy and canopy-free treatments. Positive values from 0 to 1 indicate increasingly facilitative net interactions and negative values from 0 to -1 indicate increasingly competitive net interactions. (b) Comparison of kelp health based on mean epiphytic cover of secondary laterals after the 2-month experimental period with (dark grey) and without (open) canopy ( $n = 10$  kelps per treatment). (c) Mean light illuminance recorded with and without canopy between the hours 08:00–16:00 for 14 days among the different depths. (d) Mean, maximum and minimum temperatures recorded with and without canopy over the 2 month experimental period among the different depths.

## RESULTS

### Intraspecific interactions at different depths

Intraspecific interactions between *Ecklonia radiata* and the canopy changed from net competitive to facilitative from deep to shallow along a 20 m depth gradient ( $F_{1, 7} = 7.32$ ,  $P = 0.030$ , adjusted  $R^2 = 0.442$ ; Fig. 1a). In sites deeper than 8 m, kelp growth rates were generally higher in canopy-free treatments, whereas shallower than 8 m growth rates were consistently higher within the canopy (Fig. S3). This transition in growth rate was coupled with a dramatic increase in epiphyte loading in shallow canopy-free treatments ( $73.2 \pm 9\%$  of total thallus area) compared to shallow canopy treatments ( $2.1 \pm 0.6\%$  of total thallus area) and all deep treatments (Fig. 1b; Table 1). Interestingly, after the 2 month experimental period, short-term photosynthesis and respiration rates measured in the photorespirometry chambers resulted in positive RII in five of seven sites tested, with deep sites recording the highest RII values (Fig. S4a). Large differences in interaction strength (RII) in deep sites were driven by high photosynthetic rates and low respiration in canopies compared to canopy-free treatments (Fig. S4b and c). In shallow sites, photosynthetic rates were higher than expected for canopy-free treatments, which may be attributable to the dense epiphytic cover growing on the thallus of canopy-free kelps.

Both light and temperature significantly increased with decreasing depth, however, only light was consistently modified by the canopy and represented a stress gradient per se (He & Bertness 2014; Fig. 1c and d, Table 1). Shallow canopy-free treatments had ten-fold higher mean light levels ( $14\,408 \pm 780$  Lux) than shallow canopy treatments ( $1352 \pm 329$  Lux), and four-fold higher levels than deep

**Table 1** Results of ANCOVA models comparing Growth, Epiphyte cover, Illuminance and Temperature, between canopy treatments along the depth gradient. Epiphyte cover was arcsin-square-root transformed and illuminance was log 10 transformed

		Estimate	Std. Error	<i>t</i> value	Pr(>   <i>t</i>  )	
Growth	(Intercept)	3.14928	0.361	8.722	<b>0.000</b>	
	Canopy treatment (CT)	-0.816	0.515	-1.583	0.115	
	Depth	-0.030	0.029	-1.028	0.305	
	CT:Depth	0.089	0.042	2.133	<b>0.034</b>	
	Epiphyte cover	0.156	0.062	2.530	<b>0.013</b>	
Epiphyte cover	Canopy treatment (CT)	0.784	0.091	8.589	<b>0.000</b>	
	Depth	-0.006	0.005	-1.170	0.244	
	CT:Depth	-0.051	0.008	-6.506	<b>0.000</b>	
	Illuminance (Lux)	(Intercept)	2.888	0.061	47.429	<b>0.000</b>
		Canopy treatment (CT)	1.446	0.083	17.443	<b>0.000</b>
Depth		-0.036	0.005	-7.024	<b>0.000</b>	
CT:Depth		-0.005	0.007	-0.737	0.462	
Temperature (°C)		(Intercept)	21.394	0.074	290.516	<b>0.000</b>
	Canopy treatment (CT)	0.222	0.100	2.219	<b>0.027</b>	
	Depth	-0.032	0.006	-5.193	<b>0.000</b>	
	CT:Depth	-0.015	0.008	-1.908	0.057	

Significant probabilities (<0.05) are indicated in bold.

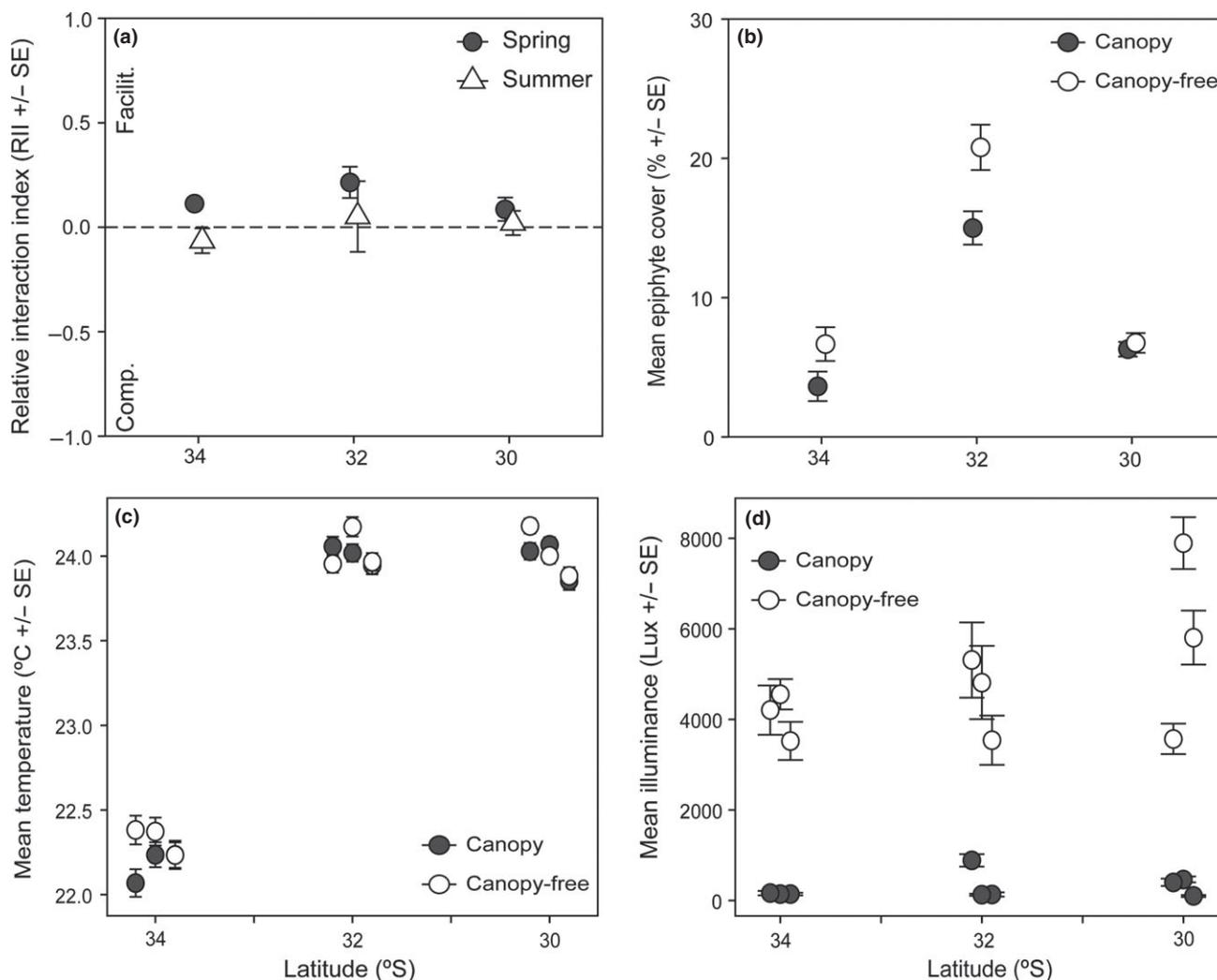
canopy-free treatments ( $3673 \pm 164$  Lux). This pattern was also consistent for maximum daily illuminance and standard deviation in daily illuminance (Fig. S5). Temperature conditions by contrast changed more uniformly with depth in both canopy and canopy-free treatments. Canopy treatments had significantly lower temperatures than canopy-free treatments, but the difference was not uniform among sites, and small (maximum difference between canopy and canopy-free treatments =  $0.29$  °C, mean difference =  $0.07 \pm 0.04$  °C) in comparison to temperature differences along the depth gradient (maximum difference between depths =  $0.73$  °C, Fig. 1d, Table 1). Furthermore, the temperature gradient was relatively benign compared to the seasonal temperature range experienced by kelps in the area ( $16$ – $23$  °C) and is therefore not considered to represent a stress gradient per se. No clear depth pattern in relative water motion (WM) was observed among the nine sites. Mean 95th percentile relative WM in both shallow and deep sites was between  $4$  and  $6$   $\text{ms}^{-2}$ , whereas the intermediate depths recorded between  $6$  and  $9$   $\text{ms}^{-2}$  (Fig. S6).

### Intraspecific interactions at different latitudes

Intraspecific interactions along the latitudinal gradient revealed predominantly positive interactions across all latitudes and did not display a significant difference in interaction strength among latitudes during either spring ( $P = 0.168$ , adjusted  $R^2 = 0.145$ ) or summer ( $P = 0.269$  adjusted  $R^2 = 0.052$ , Fig. 2a, Fig. S6). Epiphyte cover varied among latitudes, but not in accordance with the latitudinal differences in temperature (Fig. 2b and c, Table 2). Consistent with long-term profiles, summer temperatures were warmer at  $32^\circ\text{S}$  and  $30^\circ\text{S}$  than  $34^\circ\text{S}$ , but did not consistently vary between canopy and canopy-free treatments (Fig. 2c, Table 2). Despite differences in temperature among latitudes, nutrient levels remained relatively constant ( $\text{NO}_3$  concentrations =  $0.8$ – $1.5$   $\mu\text{mol L}^{-1}$ ) across all nine sites (Fig. S8). Light levels did not vary among latitudes but as in the depth experiment, mean illuminance under the canopy was significantly lower than in canopy-free treatments (Fig. 2d, Table 2). Patterns in light and temperature were consistent among mean, maximum, minimum (temperature only) and standard deviation (light only) measurements (Fig. S9 and S10).

### Generalities in intraspecific and interspecific interactions

Examining the nature of subtidal canopy interactions under low consumer pressure, across dominant species from around the world, revealed that positive and neutral interactions were widespread. Growth rates of seaweeds within the canopy were higher than canopy-free growth rates, resulting in net positive intraspecific interactions for five of the six species tested including three temperate kelp species from Australia (*E. radiata*) and South Africa (*E. maxima* and *L. pallida*), and temperate and tropical furoids (subgen. *Sargassum* sp. and *Sargassum illicifolium* respectively; Fig. 3a). *Laminaria hyperborea*, in populations from both France and Norway, was the only species to show net competitive intraspecific interactions.



**Figure 2** Nature of intraspecific interactions of *Ecklonia radiata* among nine sites along a subtidal latitudinal gradient from 34°S to 30°S. (a) Relative Interaction Index (RII) of *E. radiata* based on mean growth rates in canopy and canopy-free treatments. Positive values from 0 to 1 indicate increasingly facilitative net interactions and negative values from 0 to -1 indicate increasingly competitive net interactions. (b) Comparison of kelp health based on mean epiphyte cover of secondary laterals after the 2 month experimental period with (dark grey) and without (open) canopy ( $n = 10$  kelps per treatment). (c) Mean temperatures recorded with and without canopy over the 2 month experimental period among the different latitudes (d) Mean light illuminance recorded with and without canopy between the hours 08:00–16:00 for 14 days among the different latitudes.

Facilitation was less pronounced among interspecific interactions in seaweed canopies. Growth rates of kelp (*L. pallida*) and temperate fucoids (*Scytothalia dorycarpa*, *Arthrophykus sp.* and *Sargassum sp.*) species growing within kelp canopies displayed neutral interactions. In contrast, kelp, *E. radiata*, displayed positive interactions within temperate *Sargassum sp.* canopies while *Sargassum decurrens*, displayed net competitive interactions in tropical *Sargassum illicifolium* canopies (Fig. 3b).

## DISCUSSION

Our results revealed that positive and neutral canopy interactions are common within seaweed communities globally and that the nature of interactions can change in response to physical stress gradients in subtidal marine ecosystems where consumer pressure is low. Under benign light conditions below

8 m depth, canopy interactions were primarily competitive or neutral, whereas under shallow, high-light conditions seaweed growth became increasingly facilitated by the canopy. These findings provide support for the SGH and represent the first example of the SGH along a light-stress depth gradient that we are aware of. In contrast, interactions were neutral to positive and did not change along a latitudinal temperature gradient. While these findings do not support the SGH, they do not contradict it either, as canopy treatments were unable to consistently lower temperature conditions and alleviate temperature stress. These results were, however, consistent with interactions among nine dominant subtidal canopy species from around the world, where facilitation was the most common outcome of intraspecific interactions.

Light is a fundamental resource for plants and attenuates with depth in the ocean, limiting the vertical distribution of seaweeds (Staehr & Wernberg 2009). In the upper depths of

**Table 2** Results of ANCOVA models comparing Relative interaction index between seasons along a latitudinal gradient, Epiphyte cover, Illuminance and Temperature, between canopy treatments along a latitudinal gradient. Epiphyte cover was arcsin-square-root transformed and illuminance was square-root transformed

		Estimate	Std. Error	<i>t</i> value	Pr (>   <i>t</i>  )
RII	(Intercept)	0.956	0.935	1.022	0.325
	Season	-0.026	1.322	-0.020	0.985
	Latitude	-0.026	0.029	-0.876	0.397
	Season:Latitude	-0.003	0.041	-0.085	0.934
Epiphyte cover	(Intercept)	1.018	0.262	3.892	<b>0.000</b>
	Canopy treatment (CT)	-0.523	0.374	-1.400	0.163
	Latitude	-0.024	0.008	-2.918	<b>0.004</b>
	CT:Latitude	0.018	0.012	1.553	0.122
Illuminance (Lux)	(Intercept)	52.198	25.324	2.061	<b>0.040</b>
	Canopy treatment (CT)	103.286	35.813	2.884	<b>0.004</b>
	Latitude	1.177	0.790	1.490	0.138
	CT:Latitude	1.600	1.118	1.431	0.154
Temperature (°C)	(Intercept)	37.012	0.658	56.289	<b>0.000</b>
	Canopy treatment (CT)	-0.761	0.930	-0.819	0.413
	Latitude	0.425	0.021	20.565	<b>0.000</b>
	CT:Latitude	-0.026	0.029	-0.888	0.375

Significant probabilities (<0.05) are indicated in bold.

the ocean however, light is not a limiting resource, and high-light levels can act like a non-resource stress gradient (*c.f.* resource gradient; Maestre *et al.* 2009) whereby photosynthetically active radiation exceeds saturating levels (Altamirano *et al.* 2004) and ultra violet radiation can penetrate the water column and potentially damage kelp tissue (Wood 1987). The reduction of high-light conditions within the canopy at shallow depths appears to be fundamentally important to the health and growth of seaweed canopies. As light attenuates towards deeper depths, the dependence on neighbours reduces and interactions were observed to become neutral to competitive. Given canopies can only reduce light levels, not increase them, it would be expected that below the depths observed here, interactions would continue to become increasingly competitive as light becomes limiting towards a species vertical distribution limit. Indeed, this would explain the consistent thinning of canopies at deeper depths. Our findings are consistent with previous studies that examined recruitment patterns in response to canopy shading and observed primarily competitive interactions in depths between 10 and 18 m where canopy shading effects were similar to those observed here (Dayton *et al.* 1984; Reed & Foster 1984).

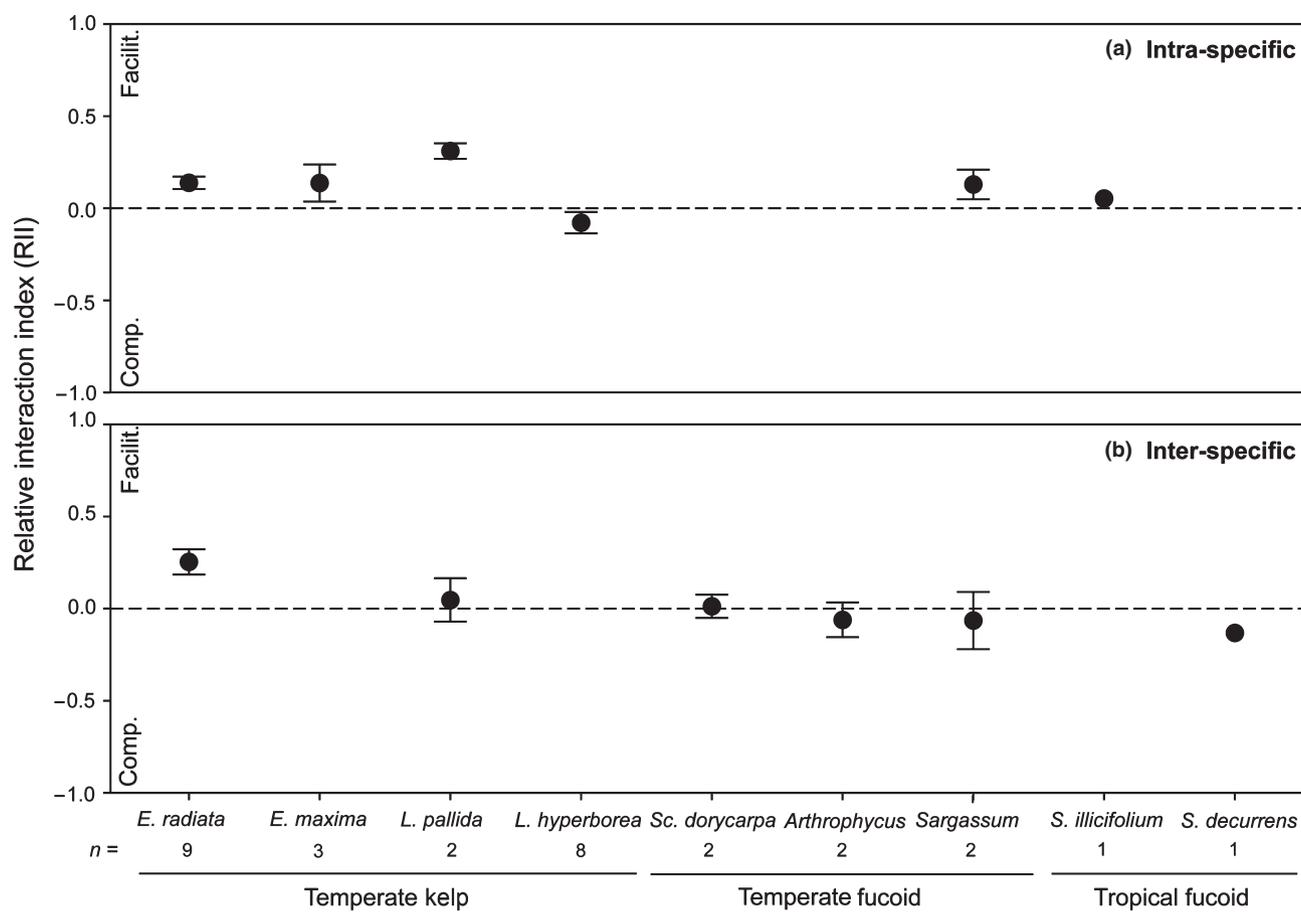
While light conditions appear to be the dominant driver of interaction patterns, other non-stressful environmental gradients (*sensu* He & Bertness 2014) were present and may have partially contributed to the nature of interactions. All depths experienced relatively high wave exposure, and no clear pattern between depth and water velocities were observed, suggesting that water velocity cannot explain the observed change in interactions with depth. Nevertheless, the presence of a canopy can reduce drag on a thallus by over 80%, compared to a solitary individual (Johnson 2001) enabling canopy

kelps to attain greater size than solitary individuals in wave-exposed areas (de Bettignies *et al.* 2013a), potentially promoting facilitation. Water movement also influences physical scour from the canopy and is likely to reduce the settlement and growth of epiphytes on kelp in the canopy. Interestingly, despite high wave exposure, temperatures changed both across the depth gradient and between canopy treatments, with slightly cooler temperatures recorded inside the canopy at some sites. Temperature differences among canopy treatments were, however, small in comparison to differences among depths and very small relative to the temperature range of *E. radiata*. Temperature is therefore unable to explain the interaction patterns observed along the depth gradient.

Temperature also did not modify interaction patterns along the latitudinal temperature stress gradient, with positive interactions prevailing at five of the six latitudinal experiments. These observations are consistent with canopy-recruit interactions observed for *Scytothalia dorycarpa* in the same region, however, differ from canopy-recruit interactions for *Sargassum* sp. and *Ecklonia radiata* which have been observed to change from competitive to facilitative in response to latitudinal temperature gradients (Wernberg *et al.* 2010; Bennett & Wernberg 2014). It remains unclear why some recruits change the nature of their interaction with the canopy along temperature gradients whereas adults do not, particularly given we observed that canopies did not modify temperatures sufficiently to reduce thermal stress. These results suggest that a combination of processes (*i.e.* light and temperature stress) can interact to determine interaction patterns in subtidal canopies (Wernberg *et al.* 2010), and that there may be ontogenetic changes in the nature of interactions within canopy species (He & Bertness 2014).

In subtidal systems elsewhere, combined effects of warm temperatures and low nutrients have been shown to influence kelp interactions. The recovery of the competitively dominant giant kelp, *Macrocystis pyrifera* following storm events depends on low-frequency cycles in nutrient and temperature conditions in response to the El Niño Southern Oscillation, which in turn influences the recruitment and abundance patterns of subordinate kelp species (Dayton *et al.* 1999). *M. pyrifera* canopies have also been adversely affected by warm temperatures and low nutrient conditions in south-eastern Australia (Johnson *et al.* 2011), but *E. radiata* canopies have not, suggesting that *E. radiata* canopies may be more tolerant of low nutrient conditions than *M. pyrifera*. High rates of canopy erosion (*c.f.* dislodgement in *M. pyrifera*) could increase nutrient supply within canopies, thereby supplementing local nutrient conditions within the canopy and promote facilitation (de Bettignies *et al.* 2013b). It remains unclear whether the small-scale clearings used in our experiment, were large enough to modify local nutrient availability, nevertheless the North American example illustrates that multiple stressors can influence the nature of interactions and may help explain the variable responses observed along the temperature gradient.

The consistent pattern of neutral to positive interactions of *Ecklonia radiata* among latitudes is in agreement with observed global interaction patterns, where positive and neutral interactions were found to be widespread among multiple



**Figure 3** Patterns in intraspecific (a) and interspecific (b) interactions between dominant subtidal benthic habitat forming species from around the world. Positive values from 0 to 1 indicate increasingly facilitative net interactions and negative values from 0 to  $-1$  indicate increasingly competitive net interactions. Numbers underneath the species name indicate the number of replicate reefs where each interaction type was measured. Sites range between 6 and 12 m depth except tropical *Sargassum* sites which were recorded in 3–4 m depth (Table S1).

subtidal canopy types that experience low consumer pressure. Among the kelp species observed, *Laminaria hyperborea* in the North Atlantic was the only species to display negative interactions with the canopy, whereas Australian and South African kelps displayed primarily positive interactions. Light availability may have contributed to these patterns as French reefs in particular were generally deeper and had lower light (due to turbidity) than canopies observed in Australia and South Africa (Staeher & Wernberg 2009). Life history traits may also help to explain the observed interaction patterns, particularly for *Sargassum* which displayed competitive interspecific interactions within the kelp canopy. *Sargassum* has a buoyant thallus that has adapted to fragment easily as a means of propagule dispersal (Schiel & Foster 2006). Fragmentation of *Sargassum* by the abrasive canopy might have contributed to the lower net growth rates of *Sargassum* within the canopy, yet may assist its dispersal and reproductive success.

In addition to contemporary environmental processes influencing interactions, the stability of canopies over evolutionary time scales may contribute to the prevalence of positive interactions, particularly in southern Australian. Modern canopy seaweeds (orders Laminariales and Fucales) evolved during

the Miocene (Silberfeld *et al.* 2010), since which time the marine climate of southern Australia has been remarkably stable (McGowran *et al.* 1997). While the diversification of Australia's marine flora was likely to have begun prior to the arrival of canopy species (Hommersand 1986), the long climatically stable history of seaweed canopies in Australia and their dense cover on reefs is likely to have favoured adaptations among canopy and understory seaweeds to thrive within canopy conditions.

Collectively, our findings highlight that canopy interactions may have an important influence on the response of subtidal habitats to physical stress. Kelp forests are often thought to be resilient to disturbance due to their fast growth rates and rapid colonisation of space (Dayton *et al.* 1984; Steneck *et al.* 2002). Our results suggest that many seaweed communities may in fact be dependent on the surrounding canopy to ameliorate stressful abiotic conditions. On shallow reefs canopy loss could result in high-light conditions that inhibit kelp recovery, providing a positive feedback mechanism that reinforces a canopy-free state. Such scenarios may become increasingly common under climate change and may help to explain subtidal community changes, as were observed following a decline in seaweed canopy cover in Western Australia in

response to an extreme marine heatwave (Wernberg *et al.* 2013). Understanding the role of other abiotic stress gradients (i.e. wave exposure) and the additive or synergistic effects of multiple stressors on interactions in the subtidal will help to improve our understanding and management of subtidal communities in the face of such global changes. The stress gradient hypothesis provides a useful concept to address these questions and has been seldom applied in subtidal ecosystems to date (Bulleri 2009). Our results reveal that neutral to positive canopy interactions are widespread throughout subtidal ecosystems globally and may be critical to reduce high-light stress in shallow subtidal marine systems. These findings suggest that canopy interactions are important in sustaining the function of many seaweed dominated systems and could therefore provide critical information about community performance across major environmental gradients and in the face of contemporary changes to the physical environment in coastal marine ecosystems.

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#### STATEMENT OF AUTHORSHIP

S.B and T.W conceived the idea and decided upon the sampling design. All authors help with data collection and logistics. S.B wrote the manuscript with contributions from all authors.

#### REFERENCES

Aberg, P. (1990). Measuring size and choosing category size for a transition matrix study of the seaweed *Ascophyllum nodosum*. *Mar. Ecol. Prog. Ser.*, 63, 281–287.

Altamirano, M.A., Murakami, A. & Kawai, H. (2004). High light stress in the kelp *Ecklonia cava*. *Aquat. Bot.*, 79, 125–135.

Anderson, M.J. (2008). Animal-sediment relationships re-visited: characterising species' distributions along an environmental gradient using canonical analysis and quantile regression splines. *J. Exp. Mar. Biol. Ecol.*, 366, 16–27.

Anderson, R.J., Carrick, P., Levitt, G.J. & Share, A. (1997). Holdfasts of adult kelp *Ecklonia maxima* provide refuges from grazing for recruitment of juvenile kelps. *Mar. Ecol. Prog. Ser.*, 159, 265–273.

Armas, C., Ordiales, R. & Pugnaire, F.I. (2004). Measuring plant interactions: a new comparative index. *Ecology*, 85, 2682–2686.

Bennett, S. & Wernberg, T. (2014). Canopy facilitates seaweed recruitment on subtidal temperate reefs. *J. Ecol.*, 102, 1462–1470.

Bennett, S., Vergés, A. & Bellwood, D. (2010). Branching coral as a macroalgal refuge in a marginal coral reef system. *Coral Reefs*, 29, 471–480.

Berec, L., Angulo, E. & Courchamp, F. (2007). Multiple Allee effects and population management. *Trends Ecol. Evol.*, 22, 185–191.

Bertness, M.D. & Callaway, R. (1994). Positive interactions in communities. *Trends Ecol. Evol.*, 9, 191–193.

de Bettignies, T., Wernberg, T., Lavery, P.S. (2013a). Size, not morphology, determines hydrodynamic performance of a kelp during peak flow. *Mar. Biol.*, 160, 843–851.

de Bettignies, T., Wernberg, T., Lavery, P.S., Vanderklift, M.A. & Moring, M.B. (2013b). Contrasting mechanisms of dislodgement and erosion contribute to production of kelp detritus. *Limnol. Oceanogr.*, 58, 1680–1688.

Bruno, J.F., Stachowicz, J.J. & Bertness, M.D. (2003). Inclusion of facilitation into ecological theory. *Trends Ecol. Evol.*, 18, 119–125.

Bulleri, F. (2009). Facilitation research in marine systems: state of the art, emerging patterns and insights for future developments. *J. Ecol.*, 97, 1121–1130.

Bulleri, F., Cristaudo, C., Alestra, T. & Benedetti-Cecchi, L. (2011). Crossing gradients of consumer pressure and physical stress on shallow rocky reefs: a test of the stress-gradient hypothesis. *J. Ecol.*, 99, 335–344.

Cavieres, L.A., Brooker, R.W., Butterfield, B.J., Cook, B.J., Kikvidze, Z., Lortie, C.J. *et al.* (2014). Facilitative plant interactions and climate simultaneously drive alpine plant diversity. *Ecol. Lett.*, 17, 193–202.

Critchley, A.T., Devisscher, P.R.M. & Nienhuis, P.H. (1990). Canopy characteristics of the brown alga *Sargassum muticum* (Fucales, Phaeophyta) in Lake Grevelingen, southwest Netherlands. *Hydrobiologia*, 204, 211–217.

Dayton, P.K., Currie, V., Gerrodette, T., Keller, B.D., Rosenthal, R. & Tresca, D.V. (1984). Patch dynamics and stability of some California kelp communities. *Ecol. Monogr.*, 54, 254–289.

Dayton, P.K., Tegner, M.J., Edwards, P.B. & Riser, K.L. (1999). Temporal and spatial scales of kelp demography: the role of oceanographic climate. *Ecol. Monogr.*, 69, 219–250.

Duarte, C.M. (1991). Seagrass depth limits. *Aquat. Bot.*, 40, 363–377.

Evans, S. & Abdo, D. (2010). A cost-effective technique for measuring relative water movement for studies of benthic organisms. *Mar. Freshw. Res.*, 61, 1327–1335.

Hay, M.E. (1986). Associational plant defenses and the maintenance of species diversity: turning competitors into accomplices. *Am. Nat.*, 128, 617–641.

He, Q. & Bertness, M.D. (2014). Extreme stresses, niches and positive species interactions along stress gradients. *Ecology*, 95, 1437–1443.

He, Q., Bertness, M.D. & Altieri, A.H. (2013). Global shifts towards positive species interactions with increasing environmental stress. *Ecol. Lett.*, 16, 695–706.

Hommersand, M. (1986). The biogeography of the South African marine red algae: a model. *Bot. Marina*, 29, 257–270.

Johnson, A.S. (2001). Drag, drafting, and mechanical interactions in canopies of the red alga *Chondrus crispus*. *Biol. Bull.*, 201, 126–135.

Johnson, C.R., Banks, S.C., Barrett, N.S., Cazassus, F., Dunstan, P.K., Edgar, G.J. *et al.* (2011). Climate change cascades: shifts in oceanography, species' ranges and subtidal marine community dynamics in eastern Tasmania. *J. Exp. Mar. Biol. Ecol.*, 400, 17–32.

Kikvidze, Z., Pugnaire, F.I., Brooker, R.W., Choler, P., Lortie, C.J., Michalet, R. *et al.* (2005). Linking patterns and processes in alpine plant communities: a global study. *Ecology*, 86, 1395–1400.

- Kitching, J.A. (1941). Studies in sublittoral ecology III Laminaria forest on the west coast of Scotland, a study of zonation in relation to wave action and illumination. *Biol. Bull.*, 80, 324–337.
- Kohler, K.E. & Gill, S.M. (2006). Coral Point Count with Excel extensions (CPCe): a Visual Basic program for the determination of coral and substrate coverage using random point count methodology. *Comput. Geosci.*, 32, 1259–1269.
- Kordas, R.L., Harley, C.D.G. & O'Connor, M.I. (2011). Community ecology in a warming world: the influence of temperature on interspecific interactions in marine systems. *J. Exp. Mar. Biol. Ecol.*, 400, 218–226.
- Long, M.H., Rheuban, J.E., Berg, P. & Ziemann, J.C. (2012). A comparison and correction of light intensity loggers to photosynthetically active radiation sensors. *Limnology & Oceanography Methods*, 10, 416–424.
- Lüning, K. (1984). Temperature tolerance and biogeography of seaweeds: the marine algal flora of Helgoland (North Sea) as an example. *Helgoländer Meeresuntersuchungen*, 38, 305–317.
- Maestre, F.T., Callaway, R.M., Valladares, F. & Lortie, C.J. (2009). Refining the stress-gradient hypothesis for competition and facilitation in plant communities. *J. Ecol.*, 97, 199–205.
- Marzinelli, E.M., Williams, S.B., Babcock, R.C., Barrett, N.S., Johnson, C.R., Jordan, A. *et al.* (2015). Large-scale geographic variation in distribution and abundance of Australian deep-water kelp forests. *PLoS ONE*, 10, e0118390.
- McGowan, B., Li, Q., Cann, J., Padley, D., McKirdy, D.M. & Shafiq, S. (1997). Biogeographic impact of the Leeuwin Current in southern Australia since the late middle Eocene. *Palaeogeogr. Palaeoclimatol. Palaeoecol.*, 136, 19–40.
- Peterson, B.J. & Heck, K.L. (2001). Positive interactions between suspension-feeding bivalves and seagrass - a facultative mutualism. *Mar. Ecol. Prog. Ser.*, 213, 143–155.
- Reed, D.C. & Foster, M.S. (1984). The effects of canopy shading on algal recruitment and growth in a giant-kelp forest. *Ecology*, 65, 937–948.
- Rodgers, K.L., Rees, T.A.V. & Shears, N.T. (In press). A novel system for measuring in situ rates of photosynthesis and respiration of kelps. *Mar. Ecol. Prog. Ser.*, doi: 10.3354/meps11273, in press.
- Schiel, D.R. & Foster, M.S. (2006). The population biology of large brown seaweeds: ecological consequences of multiphase life histories in dynamic coastal environments. *Annu. Rev. Ecol. Evol. Syst.*, 37, 343–372.
- Siddon, C.E. & Witman, J.D. (2003). Influence of chronic, low-level hydrodynamic forces on subtidal community structure. *Mar. Ecol. Prog. Ser.*, 261, 99–110.
- Silberfeld, T., Leigh, J.W., Verbruggen, H., Cruaud, C., De Reviers, B. & Rousseau, F. (2010). A multi-locus time-calibrated phylogeny of the brown algae (Heterokonta, Ochrophyta, Phaeophyceae): investigating the evolutionary nature of the “brown algal crown radiation”. *Mol. Phylogenet. Evol.*, 56, 659–674.
- Smale, D.A. & Wernberg, T. (2009). Satellite-derived SST data as a proxy for water temperature in nearshore benthic ecology. *Mar. Ecol. Prog. Ser.*, 387, 27–37.
- Stachowicz, J.J. & Hay, M.E. (1999). Reducing predation through chemically mediated camouflage: indirect effects of plant defenses on herbivores. *Ecology*, 80, 495–509.
- Staehr, P.A. & Wernberg, T. (2009). Physiological responses of *Ecklonia radiata* (Laminariales) to a latitudinal gradient in ocean temperature. *J. Phycol.*, 45, 91–99.
- Steneck, R.S., Graham, M.H., Bourque, B.J., Corbett, D., Erlandson, J.M., Estes, J.A. *et al.* (2002). Kelp forest ecosystems: biodiversity, stability, resilience and future. *Environ. Conserv.*, 29, 436–459.
- Wernberg, T., Kendrick, G.A. & Toohey, B.D. (2005). Modification of the physical environment by an *Ecklonia radiata* (Laminariales) canopy and implications for associated foliose algae. *Aquat. Ecol.*, 39, 419–430.
- Wernberg, T., Thomsen, M.S., Tuya, F., Kendrick, G.A., Staehr, P.A. & Toohey, B.D. (2010). Decreasing resilience of kelp beds along a latitudinal temperature gradient: potential implications for a warmer future. *Ecol. Lett.*, 13, 685–694.
- Wernberg, T., Smale, D.A., Tuya, F., Thomsen, M.S., Langlois, T.J., de Bettignies, T. *et al.* (2013). An extreme climatic event alters marine ecosystem structure in a global biodiversity hotspot. *Nat. Clim. Chang.*, 3, 78–82.
- Witman, J.D. (1987). Subtidal coexistence - storms, grazing, mutualism, and the zonation of kelps and mussels. *Ecol. Monogr.*, 57, 167–187.
- Wood, W. (1987). Effect of solar ultra-violet radiation on the kelp *Ecklonia radiata*. *Mar. Biol.*, 96, 143–150.

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